



Relation of wind direction and coal terminal activity patterns with air pollution burden in a community bordering a coal export terminal, Curtis Bay, Maryland, USA

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Abstract

Residents of Curtis Bay, South Baltimore, Maryland living near an open-air coal terminal have raised concerns about dark dust in their community. Coal particles have been detected in this dark dust, however, the relationship between coal terminal activity and air pollution in Curtis Bay remains unclear. This study aimed to determine how air quality varied in relation to changes in wind direction and observed coal terminal activity. We measured particulate matter (PM₁, PM_{2.5}, PM₁₀, total suspended particles) and black carbon (BC) concurrently with wind direction and visible coal terminal bulldozer activity (camera-based) at equivalent 1-minute time scales. We used conditional fixed-effects regression models to evaluate relationships between local air pollution and (1) wind direction and (2) visible bulldozer activity on coal piles. From July 5, 2022–July 16, 2023, a ten-node air monitoring network collected 2,121,793 PM and 360,325 BC 1-minute records. When downwind of the coal terminal versus not downwind, PM was 0.66 to 5.3 µg/m³ higher ($p < 0.0001$) and BC was 0.11 µg/m³ higher (95% CI=0.10, 0.12). When bulldozer activity was visible versus not visible, PM was 0.49 to 2.2 µg/m³ higher ($p < 0.001$) and BC was 0.09 µg/m³ higher (95% CI=0.08, 0.10). The associations between bulldozer activity and air pollutant concentrations were ~2.0–2.8 times higher downwind of the coal terminal versus not (all p -interaction < 0.001). PM and BC air pollution was greater in Curtis Bay when downwind of and when bulldozer activity was visible at the coal terminal, and highest when both occurred jointly. Mitigation strategies appear warranted to reduce PM and BC air pollution burden associated with coal terminal activities in Curtis Bay.

Keywords Air pollution · Particulate matter (PM) · Black carbon (BC) · Coal terminal · Coal dust · Environmental justice

Introduction

Amidst the ongoing transition to renewable energy generation, coal remains a key source of energy with global coal consumption reaching a record high in 2023 (International Energy Agency (IEA), 2023). In the United States (U.S.), there exists expansive infrastructure for coal extraction via mining and transport via rail to coal terminals for handling,

storage, and export. Coal mining activities are recognized contributors to particulate matter (PM) exposure, associated adverse health outcomes, and impacts to quality of life in both occupational (Shekarian et al. 2023; Petsonk et al. 2013; Liu and Liu 2020; Hendryx 2015; Hendryx et al. 2020b) and fenceline community (Kurth et al. 2014; Hendryx et al. 2020a; Hendryx and Entwistle 2015; Cortes-Ramirez et al. 2018) contexts.

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Compared to the state of knowledge about occupational exposure and health effects of coal mining, less is known about how coal transport, handling, and storage affect air quality, health and quality of life in neighboring fence-line communities. Recent research in a community at the fenceline of coal by rail infrastructure and an open-air coal terminal in Richmond, California linked coal transport, handling, and storage activities to local air pollution (Ostro et al. 2023a, b). This study demonstrated increases of $PM_{2.5}$ (particles of aerodynamic diameter $\leq 2.5 \mu m$) attributable to the passage of uncovered coal trains ($8.32 \mu g/m^3$ [95% CI=6.37, 10.28; $p < 0.01$]) (Ostro et al. 2023b). Although these $PM_{2.5}$ increases may be transient, their repeated periodicity was estimated to contribute to increased morbidity and mortality due to asthma and cardiovascular disease (Ostro et al. 2024). This study also found coal and petcoke (petroleum coke) dust in the $PM_{2.5}$ and coarse ($PM_{10-2.5}$) size ranges at concentrations ranging from 0.1 to $0.6 \mu g/m^3$ and 3 to $18 \mu g/m^3$, respectively, via passive aerosol sampling in the community and computer-controlled scanning electron microscopy and energy-dispersive X-ray spectroscopy (SEM-EDX) (Ostro et al. 2023a). There remains a paucity of information about air pollution burdens in communities neighboring coal terminals, which have employed a wide spectrum of mitigation strategies to control the off-site migration of coal dust and other PM air pollution into proximal communities. Questions also remain about the temporal variability of air pollution in relation to weather conditions and activity patterns at coal terminals.

In Curtis Bay, South Baltimore, Maryland (hereafter Curtis Bay), an open-air coal terminal operated by CSX Corporation is a local pollution source of longstanding community concern (Fabricant 2023). In 2023, approximately 30 million metric tons of coal were exported from the Port of Baltimore. The CSX coal terminal in Curtis Bay—one of two operating coal piers in the Port of Baltimore—has an annual throughput capacity of over 12 million metric tons (U.S. Energy Information Administration (EIA), 2024; CSX Corporation). The coal terminal is ~ 300 m from nearby homes and the Curtis Bay recreation center with local schools and businesses less than a kilometer away. For decades, Curtis Bay residents have observed and reported a daily accumulation of black dust in community spaces, at business and residential properties, and inside people's homes. In December 2021, there was an explosion at the coal terminal which showered the neighborhood with coal dust, shattered windows of nearby homes, and caused widespread panic (South Baltimore Community Land Trust 2023; Kazanjian 2022; Maryland Department of the Environment 2022).

A recent community-driven research study involving a partnership between Curtis Bay residents and several academic institutions observed coal dust (via SEM-EDX) in

settled dust collected as far as 1,235 m from the coal terminal near a local high school (Aubourg et al. 2024). Curtis Bay community residents recently documented a fugitive dust emission event (about 30 min in duration) generated by the operation of an autonomous ballast cleaner along the rail lines at the coal terminal (Shen 2024; South Baltimore Community Land Trust (SBCLT), 2024).

In response to continued local concerns and questions about coal dust exposure in the Curtis Bay community, our collaborative team of community and academic scientists aimed to collect high-resolution 1-minute-level data on air pollutants, meteorological conditions, and coal terminal activities. This study aimed to determine how changes in wind direction (downwind of coal terminal vs. not) and camera-based observations of coal terminal activity (visible bulldozer activity on coal piles vs. none) were related to changes in air pollution burden in Curtis Bay.

Materials and methods

Study setting

From 2005 to 2009, South Baltimore postal code 21226, which includes Curtis Bay, was one of the “top ten zip codes in the country for highest quantity of toxic air pollutants” (Environmental Integrity Project 2012). In the Curtis Bay area, there are around 70 stationary sources of air pollution regulated by the Maryland Department of the Environment (MDE) including—but not limited to—oil and gas processing and storage facilities, the nation's largest medical waste incinerator, a municipal wastewater treatment plant, the Baltimore City municipal landfill, and the CSX open-air coal terminal, in addition to heavy diesel truck traffic (Cassie 2024; Community of Curtis Bay Association (CCBA), 2023; Aubourg et al. 2023; Dickerson et al. 2024). The community recreation center is located about 300 m from the fenceline of the open-air coal terminal (Fig. 1). Following the 2021 explosion at the coal terminal on December 30, 2021, the Community of Curtis Bay Association (CCBA) and South Baltimore Community Land Trust (SBCLT) initiated a community-driven research collaboration with academic scientists and MDE to improve understanding of the potential impacts of the coal terminal's routine operating activities on changes in air pollution burden in Curtis Bay (South Baltimore Community Land Trust 2023; Kazanjian 2022; Maryland Department of the Environment 2022). This collaboration has been described in detail elsewhere (Aubourg et al. 2023, 2024; Dickerson et al. 2024).

Baltimore, Maryland

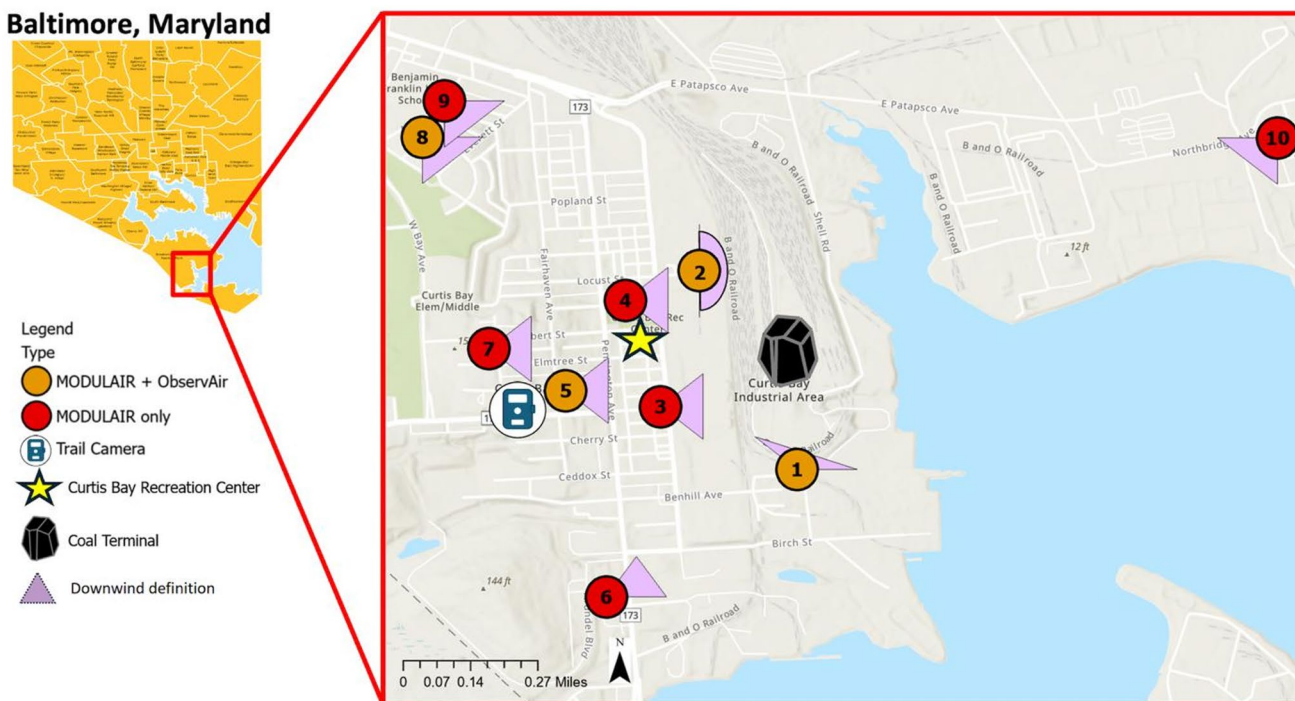


Fig. 1 Multi-pollutant air monitoring and trail camera monitoring locations, Curtis Bay, Baltimore, Maryland, U.S., 2022–2023. *Note.* Red circles indicate locations with MODULAIR units only. Orange circles

indicate locations with both MODULAIR and ObservAir units. Camera symbol indicates location of trail camera observation

Environmental and visual activity monitoring instrumentation

QuantAQ MODULAIR air monitors (QuantAQ, Inc., Somerville, MA) were deployed at ten sites to measure PM of various size fractions in units of $\mu\text{g}/\text{m}^3$: PM_{10} , $\text{PM}_{2.5}$, PM_{10} , PM_{40} or total suspended particles (TSP); gases in ppb: carbon monoxide (CO), nitric oxide (NO), nitrogen dioxide (NO_2), and ozone (O_3); and meteorological variables: temperature ($^{\circ}\text{C}$) and relative humidity (%) (Fig. 1). Additionally, each MODULAIR is connected to a sonic anemometer (Davis Instruments Corp., Hayward, CA) to record wind speed (miles per hour [mph]) and direction (degrees). At four of the ten sites (Locations 1, 2, 5, and 8), we collocated Distributed Sensing Technologies (DSTech) ObservAir air sensors, which measured black carbon (BC) in $\mu\text{g}/\text{m}^3$ (Fig. 1). Trail cameras (CamPark Electronics Co., Ltd, Hong Kong) were deployed at two locations to record coal terminal activity patterns, including visible bulldozer activity on the coal piles (Fig. 1). Footage from only one location (as shown in Fig. 1) was used in this study.

Data collection, processing & quality assurance/quality control (QA/QC)

The meteorological, trail camera, and air pollutant data were collected at 1-minute time intervals (unless noted otherwise

below). As part of our data processing and QA/QC procedure, with the exception of BC, initial data cleaning was performed at the 1-minute time scale. Then, from the 1-minute dataset, 5-, 15-, and 60-minute averages were created.

Meteorological data

Data for temperature ($^{\circ}\text{C}$), relative humidity (%), wind speed (mph), and wind direction (degrees) were collected at a 1-minute time interval at each monitoring location. Data were cleaned according to best practices, which included the removal of time periods of fog and instrument error. Wind speed and direction values were removed in cases when both wind speed and direction were equal to zero, and wind direction was removed when degrees were greater than 360° . Hourly data on solar radiation (watts per square meter [W/m^2]) were supplied by MDE from their monitoring site in Essex, Maryland, approximately 15 km (9.5 mi) from the Curtis Bay neighborhood. One minute-level observations were assigned the corresponding solar radiation value for their respective hour. At the longer time scales, wind direction was averaged based on converting the 1-minute resolution wind direction in degrees to u and v vectors, averaging these vectors to the appropriate time scales, and then converting the averaged vectors back into degrees.

The definition of a monitor being downwind of the coal terminal (Figure S2) was developed by first creating

a bearing from each respective monitoring location to the closest fenceline point of the coal terminal. The wind directions interpreted as downwind were within plus or minus 30° of this bearing. We then adjusted this angle based on the location of each air monitoring site and to reflect the unique spatial orientation aspects of the coal terminal and its operations— e.g., trajectory of rail lines within the coal terminal and docking location of coal export barges at the facility. Supplementary Results Text and Figure S2 describe the sensitivity testing to determine the downwind angle definition.

Multi-pollutant air monitoring

Particulate matter and gas data Particulate matter concentrations (PM_1 , $PM_{2.5}$, PM_{10} , TSP) were flagged as likely being impacted by fog when two conditions were met: (1) when the rolling maximum (window width=11 observations, centered on middle observation) of coarse PM ($PM_{10-2.5}$) exceeded 200 $\mu\text{g}/\text{m}^3$ AND (2) when the dew point depression (temperature – dew point temperature) was less than 3.75°C (Hagan and McClosky 2024). These heuristics are not unique to South Baltimore, therefore, we applied these criteria to PM concentrations from Curtis Bay field data and then, when possible, confirmed the presence of fog using trail cameras deployed in Curtis Bay. A limitation of this approach is the difficulty in confirming fog events during nighttime. MODULAIR units demonstrated strong agreement (as defined by EPA Non-regulatory Supplemental and Informational Monitoring [NSIM] standards) when compared to $PM_{2.5}$ and PM_{10} measurements from regulatory-grade instruments (TSP does not have an NSIM target) in Jurupa Valley, CA (Hagan 2024). Additionally, data for PM_1 and $PM_{2.5}$ may be more reliable than for larger particles because coal dust has a strong absorption component to its index of refraction and aspiration efficiencies are lower for large particles.

Similarly to the PM sensor on the MODULAIR, QuantAQ conducts laboratory-based calibrations of all gases for each unit and gas sensor for CO (linear regression), NO (hybrid regression), NO₂ (hybrid regression), and O₃ (non-linear regression). When QuantAQ collocated units with regulatory instruments in the U.S. and United Kingdom, MODULAIR units demonstrated strong agreement (as defined by EPA NSIM standards) when compared to CO, NO₂, and O₃ measurements from regulatory-grade instruments (NO does not have an NSIM target) (McClosky and Hagan 2024).

All MODULAIR units deployed in this field campaign were collocated with regulatory-grade Teledyne 640X monitors at the MDE monitoring site in Pocomoke City, Maryland from July 18 to August 7, 2023. Results from

these collocations are shown in Tables S2a-b. Based on the quality of the agreement statistics meeting EPA Air Sensor Guidelines, we did not apply additional corrections to the PM and gas data.

Black carbon We also developed data QA/QC corrections to the 2-second resolution BC data measured by the DSTech ObservAir. First, we applied a rolling 30 observation median at the 2-second time scale, whereby the rolling median of 30 consecutive 2-second observations was imputed for single 2-second time points that were identified as an outlier (defined as >1 IQR of the rolling 30 observations). Next, we instituted time-based flags, where data analysts observed erratic values during short- and long-term shutoffs. This included flagging (and removal) in two steps: (1) Flagging the following 60 min after gaps greater than 15 min between observations (including the first 60 min of data after initial startup) and (2) flagging the 5 previous observations and following 15 min after gaps greater than 2.5 min between observations. We applied this approach to each sensor. We also removed BC data on April 28 and 29, 2023 at Location 1, as the sensor at this location on these days measured concentrations in excess of 500,000 $\mu\text{g}/\text{m}^3$. Independent operators applied the flags to verify the same numbers of records removed. Finally, we removed the lowest and highest 20 BC concentration ($\mu\text{g}/\text{m}^3$) datapoints from the full combined dataset, as these were extreme values that remained after applying previous flags and removing the associated data. The DSTech ObservAir monitors were collocated with a regulatory aethalometer (Model AE33, Magee Scientific) at an MDE monitoring site. Agreement statistics have been previously documented (Aubourg et al. 2023).

Summer 2023 was a major wildfire season in Canada, resulting in region-wide poor air quality on several days. Our primary dataset excluded dates with major fire and smoke impacts in Baltimore. This list of wildfire dates was verified by MDE and includes the following days in 2023: May 12 and 31; June 1, 2, 6, 7, 8, 9, 11, 17, 18, 19, 28, 29, and 30; and July 1, 2, 17, and 18.

Trail camera-based monitoring of visible bulldozer activity at coal terminal

The trail camera was programmed to record still photographic images at 1-minute intervals. We gathered trail camera images from September 16, 2022, to June 28, 2023. Footage was reviewed to identify coal transport trains, coal pile bulldozers, coal export barges, use of the dust wet suppression system, and coal pile height. Examples of camera positioning have been previously documented (Aubourg et al. 2023). For the purposes of this analysis, visible bulldozer

activity on the coal piles was the primary activity of interest, due to its relationship with general terminal operations (e.g., adjusting piles during loading and unloading of trains) and potential to disturb coal piles and generate dust emissions.

Footage was coded “0” for no bulldozer activity observed, “1” for when bulldozer activity was observed, and “.” for times affected by complete visual obstructions. These complete visual obstructions occurred for a number of reasons including fog, lens obstruction, and periods when the camera was out of position due to high winds. When there was partial visible obstruction of the coal piles due to the presence of full tree foliage, the coding of any bulldozer activity was only based on the piles that were visible. We coded as “0” the 1-min time periods when there was no activity present on visible piles (despite potential for bulldozer activity to have been occurring on piles obstructed by tree foliage). These time periods were not coded as missing “.” because we were still able to visibly observe coal piles that were not obstructed by tree foliage. For this reason, the variable is defined as any “visible bulldozer activity” at the 1-min time scale. Only observations with a non-missing value for this variable were included in the analysis. These 1-minute level visible bulldozer activity observations were merged by date and time with air pollutant measurements from each of the ten monitoring locations. Observation totals are included in the tables of the Results section.

For “visible bulldozer activity” mean time scales longer than 1-minute (i.e., 5-, 15-, 60-min), we first created a new variable that represented the sum of all the single minute activities within each averaged time period that had visible bulldozer activity (e.g., how many minutes within a 5-minute period had visible bulldozer activity?). We then calculated the mean value across all locations of the sum of “visible bulldozer activity” variable. This led to a new variable “visible bulldozer activity above mean”, which was coded as “1” if greater than and “0” if less than or equal to the mean of the sum of visible bulldozer activity for each of the longer time scales.

Table 1 Descriptions of strata for binary downwind of the coal terminal and visible bulldozer activity variables

Binary Variable	Description of Stratum
(0, 0)	Time periods when a monitor is NOT downwind of the coal terminal and there is NOT visible bulldozer activity
(0, 1)	Time periods when a monitor is NOT downwind of the coal terminal and there IS visible bulldozer activity
(1, 0)	Time periods when a monitor is IS downwind of the coal terminal and there is NOT visible bulldozer activity
(1, 1)	Time periods when a monitor IS downwind of the coal terminal and there IS visible bulldozer activity

Inter-operator reliability was assessed between three independent operators during two points of the period of camera-based observation—October 2022 and December 2022—to assess inter-operator differences potentially related to the presence or absence of tree leaf foliage obscuring lines of sight for coding of visible bulldozer activity (Figure S3).

Data analysis

Using all meteorological, air pollutant, and trail camera observations, we created a combined dataset that merged location specific observations and trail camera observations by date and time. Analyses presented here were performed using a pooled dataset from all ten locations within the monitoring network restricted by the availability of non-missing 1-min records for the camera-based observations of any bulldozer activity at the coal terminal (n [%] missing trail camera data = 368,806 [15.7%]). We focused on the 1-min time scale to capture transient dust events with higher temporal resolution to align with community reports of frequent, short dust events. We excluded dates from the analysis that were affected by wildfires. Four of the ten monitoring locations included BC measurements (Locations 1, 2, 5 and 8 (Figure 1)). Unless otherwise noted, mean concentrations are based on a pooled mean across all locations for which data were available.

We first generated summary descriptive statistics for meteorological, air pollutant, and camera-based visible bulldozer activity variables at a 1-min time scale. Next, we calculated the mean and standard deviation (mean [SD]) air pollutant concentrations at 1-, 5-, 15-, and 60-min time intervals within each combination of strata of the binary downwind of the coal terminal variable and binary visible bulldozer activity variables (Table 1).

Relationships between (a) 1-min average wind direction and air pollutant concentrations; and (b) 1-min average visible bulldozer activity and air pollutant concentrations were evaluated by conditional fixed effects linear regression models (Allison 2009). Since the goal of this analysis was to make inferences about within-location variation of transient explanatory variables while controlling for stable characteristics of locations (both measured and unmeasured), conditional fixed effects models were more appropriate than mixed models or generalized estimating equations (Allison 2009). Conditional fixed effects models evaluate average within-location acute/transient effects by centering the repeated measures within each location (by subtracting the location-specific means) and then pooling the centered data across all locations in the study, rather than making assumptions about the distribution of between-location effects (Allison 2009). The location-specific centering ensures that potential unmeasured confounders that are location-specific

and do not vary over time do not affect the analysis results. In this design, conditional fixed effects models estimate exposure-outcome relationships by treating each location as its own control. This allows us to control for time-invariant characteristics of each location (e.g., physical structures around monitors and distance to coal terminal). We fit an interaction term between visible bulldozer activity and wind direction (downwind of coal terminal vs. not) to assess potential modification of the visible bulldozer activity–air pollution burden relationship by wind direction—during time periods when each monitor was downwind of the coal terminal compared to time periods when each monitor was not downwind of the coal terminal.

In building our conditional fixed-effects regression models we evaluated several time-varying potential confounders of the association between being downwind of the coal terminal and visible bulldozer activity and community air pollutant burden. These included temperature (°C), relative humidity (%), wind speed (mph), solar radiation (hourly W/m²), and time of day (AM/PM) because meteorological and diurnal patterns could act as confounders due to their potential to be related to both our independent (wind direction and bulldozer activity) and dependent variables (air pollutants).

For PM and BC, we also evaluated the gases CO, NO, and NO₂ as potential confounders because they indicate sources of combustion and/or vehicular traffic in Curtis Bay. These trace gases arise from combustion with urban sources dominated by vehicles (internal combustion), power plants, and industry. Carbon monoxide and NO are primary (e.g., tailpipe) emissions but NO is oxidized on time scales of minutes to NO₂; thus, NO and NO₂ are often considered together as NO_x (NO+NO₂). The average tropospheric lifetime of NO_x is hours to days. The ratio NO/NO₂ is established by ozone concentrations and in situ photochemistry again in minutes. NO is a good indicator of near-field (~1000 m) combustion. Carbon monoxide has an average atmospheric lifetime of weeks to months and is widely distributed—near sources, only small increments are seen above background levels. Cars (mostly spark ignited and gasoline-fueled) produce more CO than NO, while heavy duty trucks (mostly compression ignited and diesel-fueled) produce nearly equal amounts of CO and NO. Although we measured O₃, we did not consider it as a potential confounder of PM and BC. Ground-level O₃ is a secondary pollutant produced photochemically during atmospheric processing of NO_x (e.g., combustion emissions) and volatile organic compounds (many sources both natural and anthropogenic). Further, O₃ (or odd oxygen; NO₂ + O₃) is correlated on scales of hundreds of kilometers. Therefore, due to its strong seasonal and regional nature, O₃ was not included as a model covariate.

To evaluate the potential collinearity of covariates, we examined model variance inflation factors (VIFs) and excluded the following: relative humidity, CO, and NO₂. For conditional fixed effects regression models involving the gases as dependent variables (CO, NO, NO₂), we used the same covariate adjustment set as for PM and BC except we did not include any of the gases following an examination of collinearity.

Our final conditional fixed effects linear regression model reflects the following Eq. (1). Simplified versions appearing in the analysis involve the removal of visible bulldozer activity and the interaction term between visible bulldozer activity and downwind.

$$\begin{aligned} \text{PM_pollutant}_{it} &= \beta_1 (\text{visible_bulldozer_activity}_{it}) \\ &+ \beta_2 (\text{downwind}_{it}) \\ &+ \beta_3 (\text{temp}_{it}) + \beta_4 (\text{wind_speed}_{it}) \\ &+ \beta_5 (\text{NO}_{it}) + \beta_6 (\text{am_pm}_t) + \beta_7 (\text{solar_radiation}_t) \\ &+ \beta_7 (\text{visible_bulldozer_activity} * \text{downwind}) + \mu_i + \epsilon_{it} \end{aligned} \quad (1)$$

where: *i* indexes the observation unit of location, *t* indexes time, **β**₁, **β**₂, ..., **β**₇ are the coefficients for the independent variables, **μ**_{*i*} represents the fixed effect for each location, controlling for unobserved heterogeneity, **ε**_{*it*} is the error term. Inclusion of the fixed-effects **μ**_{*i*} implicitly does the location-specific centering, mitigating effects of unmeasured location-specific confounders (see discussion above).

To assess the consistency of our findings, conditional linear fixed effects regression models were repeated at the 1-minute time scale using two alternative definitions of the downwind angle—one wider and one narrower (Supplemental Results; Figure S2). Additionally, to assess potential variability in our results by proximity of our monitoring sites to the coal terminal, conditional linear fixed effects regression models were repeated at the 1-minute time scale, stratifying the data set by monitoring locations <0.5 km (Locations 1, 2, 3, 4) versus >0.5 km (Locations 5, 6, 7, 8, 9, 10) from the coal terminal (Supplemental Results; Tables S3–S6). This stratification by distance balanced our interest to investigate a potential role of proximity while maintaining optimal and similar sample sizes for PM and BC variables in each proximity category. To assess the consistency of results over longer time periods, we calculated stratified mean (SD) concentrations and completed conditional linear fixed-effects regression modeling at 5-, 15-, and 60-minute time scales for PM and BC parameters (Supplemental Results; Figures S4–S6; Tables S8, S10–S15). Finally, all analyses were performed for the measured gases (CO, NO, NO₂, & O₃) at the 1-, 5-, 15-, and 60-min average time scales (Supplemental Results; Figures S7–S10; Tables S7, S9, S16–S23). The primary analysis of interest was related to airborne PM and BC air pollution related to fugitive dusts, and thus analyses of

the gas variables are described in supplementary material. For example, O₃ is hypothesized to be a regional pollutant not associated with being downwind of the coal terminal or when visible bulldozer activity was occurring. All analyses were performed using R (version 4.2.3).

Results

Descriptive statistics of meteorological parameters, air pollutants, and visible bulldozer activity

Table 2 summarizes the independent meteorological and trail camera variables at the 1-minute level. Summaries are reflective of deployment times that were not always overlapping in time or season (Figure S1, Table S1) and reflect the availability of non-missing observations for ≥ 1 MODULAIR variables. Mean temperatures ranged from 9.6°C (SD=7.4) to 17.4°C (SD=8.3), with relative humidity values averaging between 50.0% (SD=18.5) and 59.3% (SD=20.1). Wind speeds demonstrated considerable variation, with means spanning 1.0 mph (SD=1.3) to 4.2 mph (SD=3.4). Solar radiation showed high variability, with mean values ranging from 122.9 W/m² (SD=201.6) to 213.2 W/m² (SD=291.5). The proportion of observations downwind of the coal terminal ranged from 10.3% to 50.0% (Table 2).

Descriptive statistics for PM (PM₁, PM_{2.5}, PM₁₀, and TSP) and BC measured at the 1-minute level are summarized in Table 3. Across all metrics, sample sizes varied by measurement, ranging from 126,630 to 291,318, reflecting differences in data availability by deployment time across locations. Mean concentrations of PM₁ ranged from 7.6 µg/m³ (SD=5.4) to 9.6 µg/m³ (SD=9.2), and PM_{2.5} exhibited similar variability, with means between 8.3 µg/m³ (SD=5.8) and 10.5 µg/m³ (SD=9.5). Variability in PM₁₀ concentrations increased with mean concentrations ranging from 16.8 µg/m³ (SD=16.0) to 39.4 µg/m³ (SD=155.2). Total suspended particulate concentrations showed high variability based on means and standard deviations, with means spanning 24.3 µg/m³ (SD=113.8) to 67.7 µg/m³ (SD=479.6).

For BC, average concentrations ranged from 0.6 µg/m³ (SD=0.5) to 1.1 µg/m³ (SD=1.8). Overall, the independent and dependent data highlight significant spatial and temporal variability in pollutant concentrations and meteorological conditions, indicative of dynamic environmental conditions as well as the varied deployment dates by location.

Visible bulldozer activity was noted in 39.9–47.3% of the data, depending on the time period of data at each location (Table 2). Daily averages for bulldozer activity ranged

Table 2 Descriptive statistics of meteorological and camera-based observations of any visible bulldozer activity at 1-minute time scale, Curtis Bay, Maryland, U.S., 2022–2023

Location	1	2	3	4	5	6	7	8	9	10	
Variable											
Temperature (°C)	n	279,422	169,163	249,332	273,630	293,032	234,965	192,655	206,228	127,688	227,279
	mean (SD)	12.7 (8.7)	11.5 (7.1)	12.3 (7.3)	13.9 (8.3)	13.0 (8.6)	13.8 (8.6)	12.8 (8.5)	9.6 (7.4)	17.4 (8.3)	13.2 (9.3)
Relative humidity (%)	n	279,422	169,163	249,332	273,630	293,032	234,965	192,655	206,228	127,688	227,279
	mean (SD)	58.1 (19.8)	50.0 (18.5)	53.0 (9.0)	53.3 (20.1)	57.5 (19.9)	51.3 (18.8)	51.8 (18.8)	59.3 (20.1)	50.4 (19.3)	58.7 (21.4)
Wind speed (mph)	n	279,879	170,615	257,435	275,350	293,463	236,251	194,067	265,568	128,132	250,393
	mean (SD)	2.3 (1.7)	4.2 (3.4)	1.9 (2.2)	1.0 (1.3)	1.0 (3.0)	1.9 (1.7)	1.9 (2.8)	2.4 (2.6)	3.2 (2.4)	2.2 (2.3)
Solar radiation (W/m ²)	n	276,460	163,539	192,205	270,042	290,286	231,121	146,384	198,879	126,542	217,491
	mean (SD)	158.0 (244.4)	147.0 (228.3)	122.9 (201.6)	150.5 (235.4)	158.2 (245.0)	165.3 (250.7)	137.6 (216.6)	127.1 (208.0)	213.2 (291.5)	162.8 (251.4)
Downwind of coal terminal (downwind = 1)	n (%)	87,872 (31.4)	59,005 (34.6)	81,341 (31.6)	102,603 (37.3)	146,458 (50.0)	24,358 (10.3)	34,768 (18.0)	59,304 (22.3)	14,960 (11.7)	47,842 (19.1)
Any visible Bulldozer Activity (yes = 1) ^a	n (%)	112,088 (40.3)	78,105 (47.3)	82,508 (42.6)	109,349 (40.3)	118,799 (40.7)	105,134 (45.2)	68,373 (46.2)	81,722 (40.7)	54,439 (42.9)	87,592 (39.9)
Average number of minutes of visible bulldozer activity per day ^a	mean (SD)	500 (376)	528 (400)	516 (385)	499 (380)	512 (382)	531 (376)	514 (395)	495 (392)	519 (363)	476 (375)
	% of day	34.8	36.6	35.8	34.7	35.6	36.9	35.7	34.4	36.0	33.1

^aDescriptive statistics include non-missing observations when ≥ 1 MODULAIR variables occurred concurrently with a trail camera observation

between 476 min (33.1% of a given 24-hour period) and 531 min (36.9% of a given 24-hour period) (Table 2).

Mean air pollutant concentrations by downwind direction and visible bulldozer activity at coal terminal

As displayed in Fig. 2, mean concentrations of all PM variables were found to be highest during periods when a monitoring site was downwind of the coal terminal and visible bulldozer activity was occurring, and lowest when the site was not downwind of the terminal and there was no visible bulldozer activity occurring. The highest vs. lowest mean (SD) for each pollutant were the following: PM_1 : 9.4 (7.1) $\mu\text{g}/\text{m}^3$ vs. 7.8 (5.9) $\mu\text{g}/\text{m}^3$; $PM_{2.5}$: 10.5 (7.1) $\mu\text{g}/\text{m}^3$ vs. 8.5 (5.9) $\mu\text{g}/\text{m}^3$; PM_{10} : 27.1 (41.7) $\mu\text{g}/\text{m}^3$ vs. 21.9 (56.2) $\mu\text{g}/\text{m}^3$; TSP: 40.2 (172.0) $\mu\text{g}/\text{m}^3$ vs. 33.4 (193.4) $\mu\text{g}/\text{m}^3$; BC: 0.93 (1.60) $\mu\text{g}/\text{m}^3$ vs. 0.67 (0.95) $\mu\text{g}/\text{m}^3$. Additionally, when bulldozer activity was not present, mean concentrations for all pollutants besides TSP were found to be elevated to their second highest level when monitoring sites were downwind of the coal terminal.

Relation between downwind direction and air pollution burden

Table 4 highlights the sample sizes, mean concentrations, and results from conditional fixed effects linear models to explore the impacts of downwind direction on 1-minute PM_1 , $PM_{2.5}$, PM_{10} , TSP, and BC concentrations. There are roughly 2,000,000 records across the ten locations for each of the PM pollutants and over 360,000 records for BC across four locations. The average concentrations (SD) for each pollutant are as follows: PM_1 : 8.3 (6.4) $\mu\text{g}/\text{m}^3$; $PM_{2.5}$: 9.1 (6.8) $\mu\text{g}/\text{m}^3$; PM_{10} : 23.5 (57.8) $\mu\text{g}/\text{m}^3$; TSP: 35.6 (215) $\mu\text{g}/\text{m}^3$, and BC: 0.76 (1.25) $\mu\text{g}/\text{m}^3$. During time periods when a monitor was downwind of the terminal, all PM and BC concentrations were higher, relative to when the monitor was not downwind of the terminal. These increases are highly statistically significant at the $p < 0.0001$ level. For an example of beta coefficient interpretation for this model, the downwind beta coefficient for $PM_{2.5}$ is 0.94, meaning the measured $PM_{2.5}$ concentration was 0.94 $\mu\text{g}/\text{m}^3$ higher when a monitor was downwind of the coal terminal versus not downwind (Table 4).

Relation between visible bulldozer activity and air pollution burden

Table 5 depicts the relation between any visible bulldozer activity at the coal terminal and air pollutant levels, stratified by wind direction at a 1-minute timescale. Sample sizes

and mean concentrations are the same as those found in Table 4. As suggested by the beta coefficients at the top of each pollutant group, when there is any visible bulldozer activity, there are increases in all pollutant concentrations on average, relative to when there is no visible bulldozer activity. For an example of beta coefficient interpretation for this model, the visible bulldozer activity beta coefficient for $PM_{2.5}$ is 0.49, meaning the measured $PM_{2.5}$ concentration was 0.49 $\mu\text{g}/\text{m}^3$ higher when visible bulldozer activity was occurring versus when not occurring (Table 5).

Joint association of downwind direction and bulldozer activity with air pollution burden

In Table 5, when there is any visible bulldozer activity at the coal terminal versus not, mean concentrations of all PM pollutants and BC are statistically significantly higher (all $p < 0.001$). Further, the bulldozer activity–air pollutant (PM_1 , $PM_{2.5}$, PM_{10} , TSP, and BC) association was larger for when an air monitor was downwind of the terminal relative to when not downwind of the terminal (all p -interaction < 0.0001). In the wind direction-stratified analysis, for all pollutants in Table 5, when the monitor was downwind of the terminal, beta coefficients were roughly two to three times higher during downwind versus not downwind stratified time periods. Figure 3 visually represents the beta coefficients from the pooled, non-downwind, and downwind results described in Sect. 3.4 and 3.5.

Discussion

This study demonstrated increased PM and BC air pollution burden in a community bordering a coal storage, handling, and export terminal which was related to time-varying patterns of wind direction and visible bulldozer activity at the terminal. Further, when both occurred jointly—i.e., during time periods when the monitoring site was downwind of the coal terminal and there was visible bulldozer activity—PM and BC air pollution burden was greatest. We used an effect measure modification analysis approach to understand evidence of statistical interaction, which demonstrated strong consistency in terms of the direction, order of magnitude, and precision of joint effect estimates across a range of transient time scales (1-, 5-, 15-, 60 min; **Supplementary Results**). The incorporation of NO as a control for local combustion and / or diesel exhaust from the bulldozers also enhanced the ability of the analysis to isolate the relationships of interest related to windblown dust.

Dust sources present particle size distributions that may be uni- or bi-modal, with tailing that extends into the finer particle sizes (Junge 1955; Brahney et al. 2024; Chalvatzaki

Table 3 Descriptive statistics for MODULAIR particulate matter (PM) and DS Tech ObservAir black carbon (BC) at the 1-minute time-weighted average, Curtis Bay, Maryland, U.S., 2022–2023

Air Pollutant	Location										
	1	2	3	4	5	6	7	8	9	10	
PM ₁ (µg/m ³)	n	277,580	164,910	193,681	271,539	291,318	219,475	138,813	200,374	126,630	218,549
	min	0.04	0.05	0.07	0.04	0.07	0.07	0.05	0.09	0.05	0.08
	mean (SD)	8.4 (6.5)	8.2 (6.2)	9.6 (9.2)	8.1 (6.1)	8.0 (6.3)	8.5 (6.0)	7.6 (5.4)	7.7 (5.3)	8.3 (6.0)	7.9 (5.6)
	max	417.0	266.0	449.0	265.1	342.0	70.0	94.0	164.6	112.3	132.2
PM _{2.5} (µg/m ³)	n	276,357	164,774	193,328	271,435	290,949	218,957	138,663	200,134	126,630	218,280
	min	0.09	0.16	0.17	0.09	0.07	0.11	0.05	0.09	0.11	0.18
	mean (SD)	9.3 (7.0)	9.0 (6.6)	10.5 (9.5)	8.7 (6.4)	8.6 (6.5)	9.5 (6.8)	8.3 (5.8)	8.4 (5.5)	9.2 (6.2)	8.8 (5.9)
	max	468.0	267.0	540.1	267.9	343.6	240.9	151.7	166.2	114.6	135.1
PM ₁₀ (µg/m ³)	n	276,357	164,774	193,328	271,435	290,949	218,957	138,663	200,134	126,630	218,280
	min	0.14	0.25	0.17	0.14	0.07	0.12	0.13	0.09	0.17	0.22
	mean (SD)	26.3 (45.5)	21.3 (38.2)	28.8 (36.5)	17.2 (40.6)	16.8 (16.0)	39.4 (155.2)	18.4 (27.9)	19.9 (19.6)	23.6 (25.5)	23.2 (24.4)
	max	5,151.0	6,814.4	2759.9	4663.7	2156.7	10524.2	5236.8	473.8	1233.6	2999.6
TSP (µg/m ³)	n	276,357	164,774	193,328	271,435	290,949	218,956	138,663	200,134	126,630	218,279
	min	0.14	0.25	0.17	0.14	0.07	0.12	0.13	0.09	0.17	0.22
	mean (SD)	39.3 (181.0)	29.4 (112.0)	43.3 (158.0)	25.1 (163.6)	24.3 (113.8)	67.7 (479.6)	28.2 (210.0)	27.9 (139.1)	38.1 (213.0)	33.7 (145.8)
	max	22,452	13,403	12,013	16,408	20,031	138,808	60,121	19,361	23,870	19,155
Black Carbon (µg/m ³)	n	137,432	137,239	--	--	35,072	--	--	185,754	--	--
	min	-10.7	-8.5	--	--	-1.3	--	--	-6.9	--	--
	mean (SD)	1.1 (1.8)	0.6 (0.7)	--	--	0.6 (0.5)	--	--	0.7 (0.9)	--	--
	max	111.8	17.0	--	--	6.2	--	--	17.2	--	--

Note: Sample size determined by available non-missing trail camera data

TSP = total suspended particles

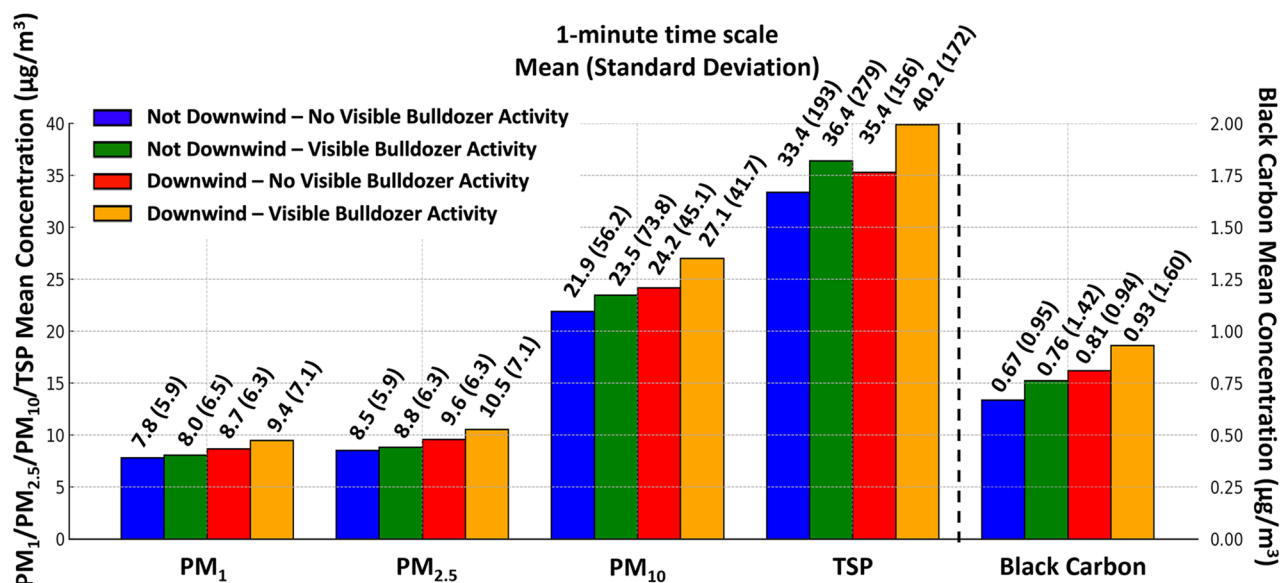


Fig. 2 One-minute average (standard deviation) concentrations of particulate matter (PM) and black carbon (BC) stratified by wind direction and visible bulldozer activity, Curtis Bay, Baltimore, Maryland, U.S., 2022–2023.

Note: Not downwind: Time periods when monitor was not downwind of coal terminal; Downwind: Time periods when monitor was downwind of coal terminal; TSP = total suspended particles

Table 4 Relation between monitor being downwind of the coal terminal and air pollutant concentration (1-min time scale), Curtis Bay, Maryland, U.S., 2022–2023

	No. of records	Mean of pollutant (SD)	Downwind Beta Coefficient ^a	95% CI	t-value	p-value
PM₁(µg/m³)						
Downwind of coal terminal (yes = 1; no = 0)	2,038,915	8.3 (6.4)	0.66	0.64, 0.68	68.8	<0.0001
PM_{2.5}(µg/m³)						
Downwind of coal terminal (yes = 1; no = 0)	2,035,542	9.1 (6.8)	0.94	0.92, 0.96	93.7	<0.0001
PM₁₀(µg/m³)						
Downwind of coal terminal (yes = 1; no = 0)	2,035,542	23.5 (57.8)	4.0	3.8, 4.2	43.6	<0.0001
TSP (µg/m³)						
Downwind of coal terminal (yes = 1; no = 0)	2,035,541	35.6 (215)	5.3	4.6, 6.0	15.1	<0.0001
Black Carbon (µg/m³)						
Downwind of coal terminal (yes = 1; no = 0)	361,863	0.76 (1.25)	0.11	0.10, 0.12	26.4	<0.0001

Note. PM = particulate matter; TSP = total suspended particles; SD = standard deviation; CI = confidence interval

^aThe beta coefficient is the increase in air pollutant (per µg/m³) when an air monitor was downwind of the coal terminal versus not downwind. Coefficients are derived from conditional fixed effects linear regression models adjusted for temperature (°C), wind speed (mph), solar radiation (W/m²), NO (ppb), and time of day (AM/PM)

et al. 2012; Seinfeld and Pandis 2006). Although the mode for the diameter of coal dust is ~10 µm, a substantial fraction of the particles fall in the respirable range under 2.5 µm. Coal dust, like wind-blown dust in general, follows a Gaussian, log-normal distribution (Junge 1955):

$$\frac{dN}{d(\log r)} = Cr^{-\beta} \quad (2)$$

where N is the number of particles, r is their radius, C is a constant, and β is a constant ~3 in the classic Junge distribution (Junge 1955). This theory has been experimentally verified for coal and is described in textbooks (Barone et

al. 2019; Seinfeld and Pandis 2006; Tessum and Raynor 2017). The large particles will fall out of the atmosphere rapidly, with a settling velocity corresponding to a transport distance of ~1 km, however, smaller particles will persist until rained out in ~1 week (Seinfeld and Pandis 2006). It is thus noteworthy that we observed consistency of associations not only for coarse PM₁₀ and TSP, but also for PM₁ and PM_{2.5}. Smaller size fraction PM, like PM₁ and PM_{2.5}, are pollutants typically associated with regional pollution, traffic, and combustion, while BC is also associated with traffic emissions. Consistency of associations with smaller PM and BC with downwind bulldozer activity may suggest that BC measurements alone or in combination with PM

Table 5 Relation between any visible bulldozer activity at the coal terminal and air pollutant levels, stratified by wind direction (1-min time scale), Curtis Bay, Maryland, U.S., 2022–2023

	No. of records	1-min mean air pollutant (SD)	Visible bulldozer activity beta coefficient ^a	95% CI	t-value	p-inter-action
Any visible coal terminal bulldozer activity (yes = 1; no = 0)						
PM₁(μg/m³)^b	2,038,915	8.3 (6.4)	0.40***	0.38, 0.42	47.3	
Wind direction (downwind of coal terminal)	579,645	9.0 (6.8)	0.74***	0.71, 0.77	45.3	< 0.0001
Wind direction (not downwind of coal terminal)	1,459,270	8.0 (6.3)	0.26***	0.24, 0.28	31.3	
PM_{2.5}(μg/m³)^b	2,035,542	9.1 (6.8)	0.49***	0.47, 0.50	54.8	
Wind direction (downwind of coal terminal)	578,530	10.0 (7.0)	0.88***	0.85, 0.92	51.9	< 0.0001
Wind direction (not downwind of coal terminal)	1,457,012	8.8 (6.6)	0.32***	0.30, 0.34	31.3	
PM₁₀(μg/m³)^b	2,035,542	23.5 (57.8)	1.3***	1.2, 1.5	16.2	
Wind direction (downwind of coal terminal)	578,530	25.5 (43.8)	2.4***	2.2, 2.6	21.1	< 0.0001
Wind direction (not downwind of coal terminal)	1,457,012	22.7 (62.5)	0.9***	0.7, 1.1	8.6	
TSP (μg/m³)^b	2,035,541	35.6 (215)	2.2***	1.6, 2.8	7.1	
Wind direction (downwind of coal terminal)	578,530	37.5 (163)	4.0***	3.1, 4.9	9.2	< 0.001
Wind direction (not downwind of coal terminal)	1,457,011	34.8 (233)	1.5***	0.7, 2.3	3.8	
Black carbon (μg/m³)^c	361,863	0.77 (1.25)	0.09***	0.08, 0.10	25.0	
Wind direction (downwind of coal terminal)	112,098	0.88 (1.32)	0.14***	0.12, 0.15	18.6	< 0.0001
Wind direction (not downwind of coal terminal)	249,765	0.72 (1.22)	0.07***	0.06, 0.08	17.3	

^aThe beta coefficient is the increase in air pollutant (per μg/m³) when any bulldozer activity was versus was not observed. Coefficients are derived from conditional fixed effects linear regression models adjusted for temperature (°C), wind speed (mph), solar radiation (W/m²), NO (ppb), and time of day (AM/PM)

^bModels include data from air monitoring locations 1 through 10

^cModels include data from air monitoring locations 1, 2, 5, and 8

*p-value < 0.05, **p-value < 0.01, ***p-value < 0.001

Note. PM = particulate matter; TSP = total suspended particles; SD = standard deviation; CI = confidence interval

may have utility for identifying black PM air pollution, like coal dust, more broadly in communities bordering coal storage, handling, and export facilities. Black carbon monitors' measurement methods select for fine mode aerosol while coal dust is generally coarse mode so future efforts should include speciated PM₁₀ or TSP.

We employed a fixed effects regression modeling approach which cannot be confounded by factors that do not vary at the same time scale as the temporal scale of the exposure-response being investigated. The results of this regression modeling approach for PM₁, PM_{2.5}, PM₁₀, TSP and BC support the hypothesis that mean concentrations of these air pollutants are higher when the winds blow toward monitoring locations downwind from the coal terminal. The statistical methodology adopted has certain limitations. When only considering downwind wind direction as the exposure, there is a chance of potential confounding from another source of pollution coming from the same direction as the coal terminal. This confounding risk is mitigated to a large extent when using the visible bulldozer activity data as the exposure as it is unlikely that there is another temporal confounder that is highly concurrent with bulldozer activity, particularly as we control for local combustion and / or diesel exhaust from the bulldozer activity. The assumption of linearity in the fixed effect models used can also be

violated. The current analysis also only offers an average effect size based on pooled data from all ten air monitoring sites. Separate site-specific analysis can reveal variability in these effect sizes and how they relate to distance from the coal terminal.

Our focus on transient exposure-response time scales reflects the community lived experience involving reports of frequent short term dust events that recur frequently (analogous to our 1-min, 5-min, and 15-min time scales) as well as longer events that have been documented (analogous to our 60-min time scale). Curtis Bay community members recently documented an example of a longer high visibility fugitive dust event at the coal terminal which lasted for approximately 30 min (Shen 2024; South Baltimore Community Land Trust (SBCLT), 2024). The consistency of our exposure-response associations across all times scales considered (1-min, 5-min, 15-min, and 60-min) for both explanatory variables (downwind of the coal terminal and visible bulldozer activity) and all PM and BC air pollutants (Figs. 3, S4-S6; Table S8) is coherent with neighbors' reports about the nature of fugitive dust events.

To our knowledge, this study is the first to investigate the relation between wind direction and coal terminal activity upon the air pollution burden in a proximal fenceline community. This study has some limitations including TSP

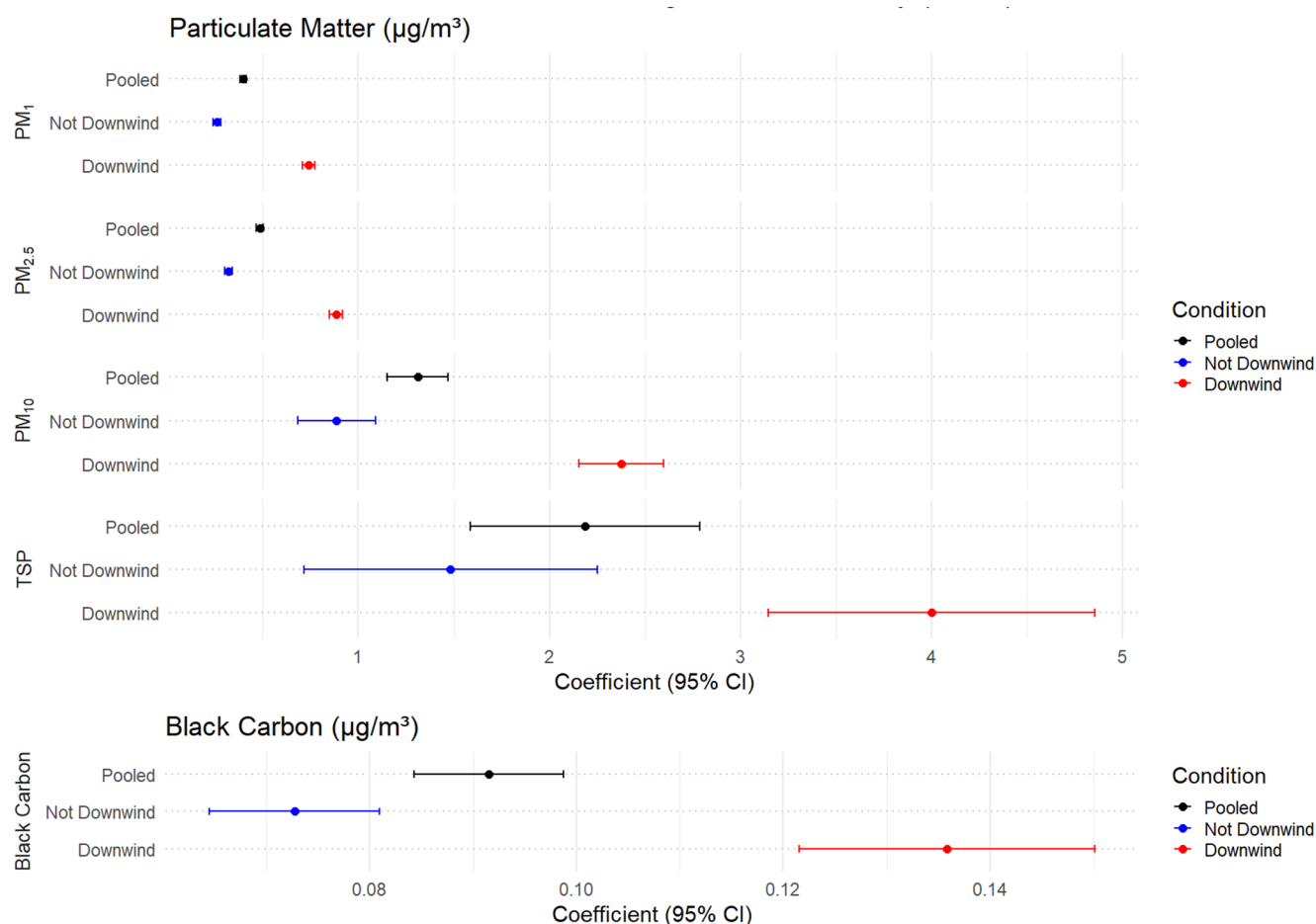


Fig. 3 Forest plot showing magnitude of beta coefficient and 95% confidence interval (CI) estimates of the relation between any visible bulldozer activity at the coal terminal and changes in PM_{10} , $\text{PM}_{2.5}$, PM_{10} , total suspended particles (TSP) and black carbon (BC) stratified by time periods that were downwind of the coal terminal versus not at 1-minute average time intervals, Curtis Bay, Baltimore, Maryland, U.S., 2022–2023. *Note.* Binary any visible bulldozer activity versus

none was defined as: If a 1-minute time period had any visible bulldozer activity=1; if a 1-minute time period did not have any visible bulldozer activity=0. All p -interaction < 0.001 for PM_{10} , $\text{PM}_{2.5}$, and PM_{10} , TSP, and BC. Coefficient = Beta coefficient of conditional fixed effects linear regression model for the relation between any visible bulldozer activity at the coal terminal and air pollutant levels

measurements by moderate-cost sensors do not have a convenient reference comparator for validation and calibration, unlike $\text{PM}_{2.5}$, PM_{10} , and BC. Observed variation in TSP values with large standard deviations may be attributed to instrument error or natural variability due to the higher settling velocity of larger particles resulting in spatial heterogeneity on scales of ~100 m. In the coding of trail camera data, tree foliage inhibited the identification of bulldozer activity on certain coal piles during certain times of year. During these times, if there was no bulldozer activity at visible coal piles, the observation was coded as no activity “0” rather than missing “.”. We hypothesize that this could have led to an underestimate of the true extent of bulldozer activity that was occurring on the coal piles during these tree foliage-impacted times of year. However, this potential misclassification of bulldozer activity was likely non-differential with respect to the outcome (air pollution).

Our study also has several notable strengths. Our research questions were informed by the concerns, observations, and experiences of Curtis Bay community members who have lived proximal to the coal terminal for decades. This guided our analytic approach involving conditional fixed-effects regression modeling of multi-pollutant data. Thereby, we controlled for time-invariant characteristics of each location, such as proximity to pollution sources or local infrastructure, which could otherwise confound our observed association. By focusing on within-monitor variation in air pollutants over time, fixed-effects models isolated the effects of wind direction and visible bulldozer activity while removing the potential influence of location-specific non-time varying factors. Another strength is the use of the MODULAIR multi-pollutant air sensor which monitors dust in multiple particle size bins. This may be used for quality control to improve confidence in results— i.e., temporal

patterns observed in multiple, independent channels are more realistic in mixed urban/industrial environments than a unique signal in a discrete particle size bin.

Future directions for this line of research could extend the conditional fixed effects regression modeling approach to other time-resolved activity patterns at the coal terminal. While our analysis focused on visible bulldozer activity due to its association with general terminal operations, future analyses can be extended to code time-resolved activity patterns of coal transport by trains, coal ship / barge activity, coal pile height, and use of the dust wet suppression system. Additionally, the conditional fixed-effects regression modeling approach used assumes a linear relationship, and future studies could explore other non-linear approaches, assess and account for potential unmeasured temporal confounding and lags. Our preliminary analysis of how air quality burden varies with proximity of the measurement sites to the coal terminal could be expanded to investigate location-specific and other spatial patterns. Further, there are opportunities for source attribution analyses to identify combinations of air pollutants that are associated with coal terminal emissions.

Short-term exposure to PM has been increasingly linked to adverse health outcomes, with more research still needed (Gutiérrez-Avila et al. 2023; Orellano et al. 2020). Fine PM (PM_{10} and $PM_{2.5}$) are of particular concern due to their ability to penetrate deep into the respiratory tract and enter the bloodstream, exacerbating cardiovascular and respiratory conditions. Acute exposure to PM_{10} has been associated with increased hospital admissions for asthma, bronchitis, and chronic obstructive pulmonary disease (COPD), as well as heightened risks of stroke and cardiac events (Alesandrini et al., 2016, Orellano et al. 2020). These impacts are particularly pronounced in vulnerable populations such as children, the elderly, and individuals with pre-existing health conditions. Both long-term and acute exposure to BC have been linked to morbidity and mortality due to cardiovascular and respiratory disease (Song et al. 2022; Luben et al. 2017; Hvidtfeldt et al. 2019; Ostro et al. 2015; Janssen and Joint 2012), in addition to marked impacts upon fetal, pediatric, and maternal health outcomes (Paunescu et al. 2019; Goriainova et al. 2022). Residents in communities of color and low wealth are also disproportionately exposed to PM, BC, and other forms of air pollution, experience elevated impacts of this pollution, and are often overburdened by the cumulative impacts of other chemical and non-chemical stressors. Higher $PM_{2.5}$ exposure was also observed for Hispanic and Black residents in San Francisco, CA living near uncovered coal by rail transport infrastructure (Ostro et al. 2024).

Community observations, actions taken by state/local governments, and the results of this research suggest

additional measures are warranted to end or mitigate fugitive dust emissions related to coal terminal operations and neighborhood health and quality of life impacts associated with such emissions. Currently, the facility maintains a wet dust suppression system and has a partial porous wind fence at the southern edge of the coal terminal. When restricting our analysis to the one location within the network on the opposite side of the wind fence, our results were generally consistent with the analysis that pooled data across all locations. This suggests that off-site community air pollution burden is not reduced by wind screen netting mitigation strategies. A range of additional strategies have been submitted by community members. These include, but are not limited to: (1) denial of the application for the renewal of its operating permit based on a record of complaints and violations; (2) full enclosure of the coal piles paired with enforceable limits on PM from the terminal; (3) the imposition of fees on a per-ton basis of coal transferred through the terminal to mitigate negative externalities; and (4) transitioning to the storage and handling of a safer material at the facility.

Conclusions

The findings of this study demonstrated higher PM and BC air pollution burden in Curtis Bay when downwind of the coal terminal and when there was visible bulldozer activity—highest air pollution burdens were observed when both occurred jointly. Results are consistent with the lived experiences of Curtis Bay community members who have documented ongoing, repeated patterns of black dust accumulation at and around their homes alongside recordings of transient fugitive black dust plumes at the coal terminal. Further context for our findings is provided by a recent study demonstrating the detection of coal dust in the black dust that accumulates in residential and community spaces of the Curtis Bay neighborhood (Aubourg et al. 2024). Overall, our findings in context with complementary research and community experiences contribute to a coherence of data across varied analytical and observational platforms that respond to residents' questions and concerns with dark dust accumulation in Curtis Bay. Compared to both Baltimore City and the State of Maryland, the Curtis Bay community is overburdened by cumulative impacts of economic hardship, high vacancy rates, historical pollution burden, and health disparities such as higher mortality due to chronic obstructive pulmonary disease and lung cancer (Baltimore City Health Department 2017). Strategies to mitigate the observed air pollution burdens in the Curtis Bay community appear warranted, including those that could prevent

windblown and mechanical dusts that might be associated with operating practices at the coal terminal.

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Data availability De-identified data will be made available upon request.

Declarations

Ethics approval and consent to participate This study was reviewed and determined to be not human subjects research (NHSR) by the Johns Hopkins Bloomberg School of Public Health Institutional Review Board (IRB).

Consent for publication Not applicable.

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